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我国温室大棚邻苯二甲酸酯(PAEs)污染及综合控制技术研究进展

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摘要:文章综述了我国温室大棚邻苯二甲酸酯(PAEs)的主要来源,包括农膜、化肥农药和污水灌溉等;收集了珠三角、长三角、环渤海和东北地区温室大棚种植系统中土壤和蔬菜PAEs污染调查研究的数据,并与露天种植系统中PAEs含量进行对比统计分析;评价了不同地区的不同年龄人群对蔬菜PAEs的日摄入量和健康风险;总结了温室大棚PAEs污染控制技术,为控制温室大棚蔬菜PAEs污染提供了科学依据。

关键词:温室大棚;邻苯二甲酸酯;污染状况;人体健康风险评价;污染控制

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Current status of phthalate acid esters(PAEs) in greenhouses in China and comprehensive control technology:A review

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Abstract: Phthalate acid esters(PAEs) are widely used as plasticizers and constitute one of the most frequently detected organic environmental contaminants. With the deterioration of the eco-environment in China during the past three decades, many studies on the occurrence of PAEs in greenhouse soil-crop systems and their risk evaluations have been conducted; this allows us to carry out a fairly comprehensive assessment of PAE contamination at a nation-wide scale. The sources of PAEs in agricultural soil mainly include agricultural film, chemical fertilizer, and sewage irrigation. We collected published data of PAE levels in soil-crop systems(greenhouse vegetable and open field planting systems) from the Pearl River delta, Yangtze River delta, Bohai rim, and northeast China. The concentrations of dibutyl phthalate(DBP) and di(2-ethylhexyl) phthalate(DEHP) in vegetables in the Yangtze River delta were higher than those in the Pearl River delta and Bohai rim, while the concentrations of DBP and DEHP in soils in northeast China were lower than those in other regions. The daily intake of PAEs was evaluated for different age groups from different regions of China, suggesting potential health risks. The control strategies for greenhouse PAE pollution are summarized and offer a scientific basis for the control of PAEs in greenhouses.

Keywords:greenhouse;phthalate acid esters;contamination status;human health risk assessment;pollution control

邻苯二甲酸酯(Phthalate acid esters, PAEs)通常作为增塑剂被广泛用于食品包装袋、农用塑料薄膜,以增强塑料产品的弹性和柔韧性^[1],并作为添加剂用

于肥料、油漆、个人护理产品及杀虫剂等^[2]。PAEs全球产量高达3亿t·a⁻¹^[3],它们与塑料之间以非共价键形式结合^[4],在生产、使用和处理过程中容易被释放

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进入环境。由于PAEs的广泛使用^[4],导致其在土壤、水、空气、污泥等环境介质和生物体内普遍检出^[5]。PAEs是典型的内分泌干扰物,具有潜在的生殖毒性和“三致”效应^[6-7],因此,美国环境保护局(USEPA)将邻苯二甲酸二甲酯(DMP)、邻苯二甲酸二乙酯(DEP)、邻苯二甲酸二丁酯(DBP)、邻苯二甲酸二正辛酯(DnOP)、邻苯二甲酸二(2-乙基)己酯(DEHP)、邻苯二甲酸丁苄酯(BBP)等6种PAEs化合物列为优先控制有机污染物^[8]。我国也将其中3种PAEs化合物(DMP、DBP、DOP)确定为环境优控污染物^[9]。

温室大棚技术能够突破季节限制实现蔬菜的速生、优质和高产,是满足城市居民对蔬菜消费需求的重要保障和现代化农业生产的手段^[10-11]。温室大棚使用大量塑料棚膜,其中的PAEs可达20%~60%^[5]。在生产过程中,PAEs会逐渐释放进入大棚土壤及其空气中,易被蔬菜吸收,并累积在可食部位。长期食用含有PAEs的农产品会导致低剂量暴露,对人体产生一定健康风险^[12]。温室大棚常年处于封闭或半封闭状态,具有高温、高湿、高蒸发量和无雨水淋洗等特点,长期大量施用化肥导致土壤PAEs污染,进一步增加了大棚蔬菜PAEs残留。北京、新疆、辽宁、吉林和福建5个农膜使用区土壤PAEs含量>10 kg·hm⁻²·a⁻¹^[13];黑龙江、辽宁、山东、南京等地的蔬菜大棚土壤也受到不同程度的PAEs污染^[14-15],大棚蔬菜中PAEs含量达到每千克几毫克至几十毫克^[16-17]。随着温室大棚栽培面积的不断扩大,农膜使用量逐年增加,其引发的农产品安全问题和土壤环境问题日益突出,但目前对于我国温室大棚土壤-蔬菜系统PAEs的污染状况还缺乏整体了解。

目前关于设施农田土壤以及蔬菜PAEs污染研究已有一些报道,但关于我国温室大棚PAEs污染状况及综合控制技术的综述论文还鲜见报道。为系统了解我国温室大棚中的PAEs污染现状及其所引发的潜在人体健康风险,本研究检索了近几年Web of Science(WOS)和中国知识基础设施工程(CNKI)数据库中的文献共33篇,对珠三角($n_{soil}=691$, $n_{vegetable}=346$)、长三角($n_{soil}=360$, $n_{vegetable}=454$)、环渤海($n_{soil}=403$, $n_{vegetable}=120$)、东北($n_{soil}=68$)等地区温室大棚和露天种植系统中土壤、蔬菜PAEs污染调查数据进行了统计分析(表1),总结了我国温室大棚土壤和蔬菜中PAEs污染状况,评价了蔬菜中PAEs的人体日摄入量和健康风险,提出了控制温室大棚PAEs污染的综合技术,以期为温室大棚的安全生产和保障农产品安全提供

科学依据。研究发现温室大棚和蔬菜中DBP和DEHP浓度和检出频率最高,因此本文以DBP和DEHP作为评价温室大棚土壤-蔬菜系统中ΣPAEs污染的代表性污染物。

1 温室大棚系统PAEs来源

1.1 农膜使用

随着我国城市化迅速发展,温室蔬菜生产规模不断增大,农膜覆盖面积居世界第一位^[18]。至2016年,用于农业生产的塑料薄膜数量达260万t,地膜使用量达到147万t,农膜覆盖面积达1840万hm²。我国农膜使用量在不同区域和省份有较大差异,地膜覆盖主要分布东北、华北、华东、华中、西南和西北地区,其中黑龙江、河北、山东、河南、四川等省份农膜使用量高于其他省区^[49]。新疆、山东和甘肃等蔬菜和棉花产区的农膜使用和地膜覆盖面积居全国前3位。新疆、甘肃等西北部地区属干旱半干旱地区,农业生产中会大量使用农膜保水,其中新疆棉田土壤中PAEs含量最高,达到11.2~1232 mg·kg⁻¹^[50]。然而,关于我国西北地区大棚系统中土壤和蔬菜PAEs污染鲜有报道,需要加强该地区农田(大棚)土壤-蔬菜系统PAEs状况的调查研究。由于农膜使用量大,回收成本高,全国地膜平均残留率为19.7%,每年有大量农膜残留在土壤中,并不断积累^[51],导致土壤和蔬菜具有较高的PAEs污染风险。有研究报道,温室大棚土壤中PAEs浓度是露天土壤的2.5~3倍^[52],而温室大棚蔬菜中PAEs含量是露天种植蔬菜的3~6倍^[35,53]。

温室大棚土壤蔬菜PAEs污染与农膜中PAEs含量、膜厚度、大棚高度、棚龄和地膜颜色有关,如农膜中DEHP含量越高、棚膜越厚、大棚高度越低、棚龄越小,芥菜、白菜、芹菜、菠菜、卷心菜、管叶、生菜、大蒜和苋菜等9种蔬菜中DEHP含量越高^[17]。白色农膜中主要PAE成分为DEHP,其ΣPAEs(119.39 mg·kg⁻¹)要显著高于黑色农膜(50.84 mg·kg⁻¹)^[36]。对于长期使用黑色塑料膜覆盖的土壤,PAEs浓度也非常高,可能由于黑色地膜更容易吸收热量,地膜温度升高,加速了PAEs的释放^[33]。通过4种不同种植模式的温室大棚土壤PAEs浓度发现,温室大棚土壤中PAEs浓度还与其种植模式有关^[36]。温室大棚+小棚+地膜+蔬菜模式与温室大棚+地膜/或仅地膜+蔬菜模式中PAEs含量最高,约为3.18 mg·kg⁻¹,其次是温室大棚+地膜+蔬菜模式(1.88 mg·kg⁻¹)和温室棚膜覆盖/不覆盖交替+蔬菜模式(0.81 mg·kg⁻¹)。

表1 我国不同地区土壤和蔬菜中PAEs的含量($\text{mg}\cdot\text{kg}^{-1}$ DW)Table 1 The content of PAEs in soil and vegetable in different regions of China ($\text{mg}\cdot\text{kg}^{-1}$ DW)

| 位置 Location | 类型 Types | 蔬菜中DBP含量 | | | 蔬菜中DEHP含量 | | | 土壤中DBP含量 | | | 土壤中DEHP含量 | | | 参考文献 Ref. |
|-----------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|--------------|-------------|-------------|--------------|--------------|
| | | 平均值 Mean | 标准差 S.D. | 范围 Range | 平均值 Mean | 标准差 S.D. | 范围 Range | 平均值 Mean | 标准差 S.D. | 范围 Range | 平均值 Mean | 标准差 S.D. | 范围 Range | |
| 珠三角 地区 | 露天种植 | 0.002 | 0.000 2 | 0.002~0.023 | 0.291 | 0.146 | 0.006~0.465 | 0.002 | 0.001 | 0.001~0.004 | 0.021 | 0.017 | 0.001~0.057 | [18] |
| | 露天种植 | — | — | — | — | — | — | 0.342 | 0.423 | 0.009~2.740 | 1.760 | 1.590 | 0.107~29.400 | [19] |
| | 露天种植 | — | — | — | — | — | — | 9.230 | — | nd~20.600 | 11.000 | — | 2.820~25.100 | [20] |
| | 露天种植 | 0.562 | 0.376 | 0.162~1.790 | 1.920 | 1.380 | 0.046~6.100 | 0.164 | — | 0.036~0.282 | 0.410 | — | 0.074~1.470 | [21] |
| | 露天种植 | 0.874 | 0.801 | nd~9.920 | 0.718 | 1.140 | nd~38.100 | — | — | — | — | — | — | [22] |
| | 露天种植 | — | — | — | — | — | — | 0.259 | 0.162 | nd~17.500 | 0.148 | 0.080 | nd~6.480 | [23] |
| | 露天种植 | 0.250 | 0.062 | nd~0.680 | 0.120 | 0.056 | nd~0.740 | 0.673 | 0.092 | nd~1.720 | 0.183 | 0.103 | nd~1.860 | [24] |
| | 露天种植 | 0.553 | 0.030 | 0.110~1.100 | 0.668 | 0.116 | 0.180~2.760 | 0.775 | 0.136 | 0.370~1.480 | 1.660 | 0.575 | 0.270~4.970 | [25] |
| 长三角 地区 | 露天种植 | — | — | — | — | — | — | 0.282 | — | nd~1.770 | 0.140 | — | nd~1.390 | [26] |
| | 露天种植 | 6.840 | 7.090 | nd~17.100 | 0.428 | 0.543 | 0.041~0.643 | 0.409 | 0.559 | nd~7.650 | 0.146 | 0.126 | 0.001~4.200 | [27] |
| | 露天种植 | 0.578 | 0.852 | nd~1.840 | 0.147 | 0.195 | nd~0.540 | 0.980 | 0.495 | nd~1.330 | 0.710 | 0.994 | nd~2.180 | [28] |
| | 露天种植 | 3.900 | 3.240 | nd~9.270 | 3.020 | 2.000 | nd~7.090 | 1.670 | — | — | 1.940 | — | — | [29] |
| | 露天种植 | 1.890 | — | 0.777~2.220 | 0.610 | — | 0.311~0.818 | 0.370 | — | — | 0.770 | — | — | [30] |
| | 设施农业 | 0.021 | 0.006 | 0.009~0.057 | 0.023 | 0.006 | 0.015~0.056 | 0.065 | 0.022 | 0.016~0.187 | 0.024 | 0.010 | 0.003~0.118 | [31] |
| | 设施农业 | 0.092 | — | nd~0.659 | 0.796 | — | 0.029~5.520 | 0.100 | — | nd~0.720 | 0.450 | — | nd~2.250 | [32] |
| | 设施农业 | — | — | — | — | — | — | 0.210 | — | 0.140~0.350 | 1.480 | — | 0.810~2.200 | [33] |
| 环渤海 地区 | 设施农业 | 0.800 | — | 0.130~1.810 | 1.370 | — | 0.120~5.820 | 0.170 | — | nd~2.080 | 1.840 | — | 0.240~4.180 | [34] |
| | 露天种植 | 3.020 | 4.580 | 0.210~8.300 | 3.450 | 5.010 | 0.543~9.230 | — | — | — | — | — | — | [35] |
| | 露天种植 | 0.983 | 0.460 | 0.660~1.510 | 0.523 | 0.206 | 0.330~0.740 | — | — | 7.790~15.100 | — | — | 5.140~13.700 | [36] |
| | 设施农业 | — | — | — | — | — | — | 0.162 | 0.152 | 0.054~0.269 | 0.101 | 0.040 | 0.073~0.129 | [37] |
| | 设施农业 | 0.400 | — | — | 0.460 | — | — | 0.435 | — | — | 0.671 | — | — | [38] |
| | 设施农业 | — | — | — | — | — | — | 1.470 | — | 0.016~15.700 | 1.470 | — | 0.073~5.320 | [16] |
| | 露天种植 | — | — | — | — | — | — | 0.165 | — | — | 0.046 | — | — | [39] |
| | 设施农业 | 0.498 | — | 0.016~2.950 | 0.549 | — | 0.049~2.630 | — | — | — | — | — | — | [40] |
| 东北及 其他地 区 | 设施农业 | — | — | — | — | — | — | 0.425 | — | — | 0.369 | — | — | [41] |
| | 露天种植 | 1.181 | 0.319 | 0.010~3.330 | 1.600 | 0.313 | 0.110~4.500 | 3.070 | — | 0.550~10.800 | 1.080 | — | 0.420~7.060 | [42] |
| | 设施农业 | — | — | — | — | — | — | 1.140 | — | 0.340~1.660 | 0.470 | — | 0.220~0.740 | [43] |
| | 设施农业 | 0.440 | — | — | 0.380 | — | — | 0.500 | — | — | 0.550 | — | — | [44] |
| | 露天种植 | — | — | — | — | — | — | — | — | nd~0.112 | — | — | nd~0.274 | [47] |
| 露天种植 | — | — | — | — | — | — | — | — | — | 0.140~0.350 | — | — | 0.810~2.200 | [48] |

1.2 粪肥施用

常用肥料PAEs含量为 $0.01\sim3.00\text{ mg}\cdot\text{kg}^{-1}$ ^[48,54]。据统计,2015年中国化肥消费量达到6020万t,是1990年的两倍多(2590万t);此外,鸡粪、猪粪、牛粪、鸭粪等有机肥料中也被检测出含有PAEs,其含量为 $2.24\sim6.84\text{ mg}\cdot\text{kg}^{-1}$ ^[48],高于化肥中的PAEs浓度。施用的大量粪肥也是温室大棚种植系统PAEs的重要来源。因此,化肥、有机肥的大量使

用将增加PAEs引入,增大温室大棚土壤和蔬菜的生态风险。

1.3 灌溉水源与灌溉方式

灌溉水源和方式是影响温室大棚种植系统PAEs污染浓度的又一重要因素。温室大棚灌溉水源一般来自地表水、地下水和自来水^[54],然而这些水源中常检测出不同含量的PAEs^[55~56]。我国每升地表水中PAEs检出浓度为几微克到十几微克^[57~58]。此外,污水

作为灌溉水源,既可以增加土壤肥效,又可作为干旱/半干旱地区大棚种植的重要水源之一^[19]。但目前我国污灌水质缺乏有效监管,大量未经处理的污水直接应用于农田灌溉,造成农田PAEs污染。如广州某污水灌溉农业土壤中ΣPAEs含量高达33.6 mg·kg⁻¹,其中DEHP含量为29.4 mg·kg⁻¹;长期污水灌溉的天津郊区农业土壤中,DEHP和DBP浓度也相对较高^[59]。灌溉方式也会影响农田土壤中PAEs浓度与分布,如滴灌土壤中PAEs浓度要低于漫灌土壤,其中滴灌处理土壤PAEs主要分布在0~10 cm的表层土,漫灌处理土壤PAEs主要分布于10 cm以下的土层^[60]。

2 温室大棚系统PAEs污染现状与日摄入量风险评价

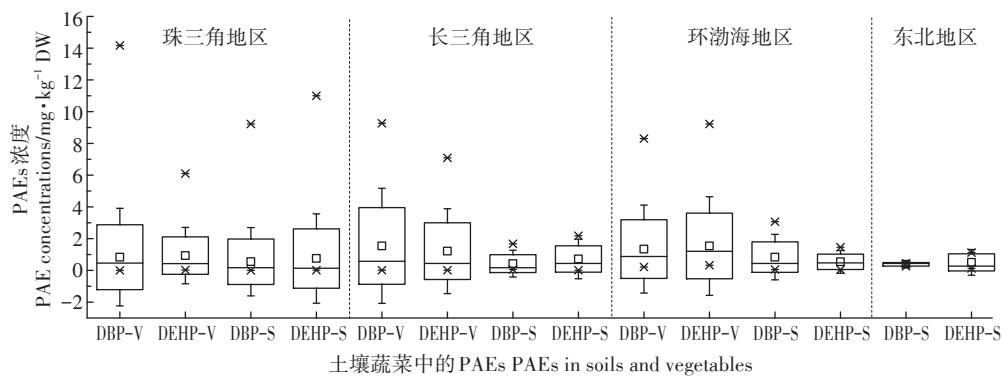
2.1 我国土壤和蔬菜PAEs含量的地区差异

珠三角、长三角、环渤海和东北地区是中国工业和经济发达、人口众多、蔬菜消费量大的几个地区,其农膜使用量也位居全国前列,其中环渤海地区约100万hm²、长江三角洲约40万hm²、东北地区约32万hm²和珠江三角洲地区约21万hm²^[61]。珠三角和长三角地区土壤中PAEs含量:DEHP>DBP,环渤海地区土壤中PAEs含量:DBP>DEHP,而珠三角和长三角地区蔬菜中PAEs含量:DBP>DEHP,环渤海地区蔬菜中PAEs含量:DEHP>DBP。土壤DBP含量最高的是环渤海地区(0.837 mg·kg⁻¹),而DEHP含量最高的是珠三角地区(0.794 mg·kg⁻¹);长三角地区蔬菜中DBP和DEHP含量较珠三角和环渤海地区要高。调查区域内蔬菜中PAEs含量普遍高于土壤PAEs含量,其中东北地区土壤中的PAEs含量低于其他地区(图1)。经

统计分析,不同地区蔬菜、土壤PAEs含量与农膜使用量相关系数(R^2)分别为0.975($P<0.01$)和0.457($P<0.05$),表明这可能与该4个地区的农膜使用量有密切的关系。目前,我国尚未制定农田土壤PAEs控制标准。美国土壤PAEs化合物控制标准DBP为0.081 mg·kg⁻¹、DEHP为4.35 mg·kg⁻¹。参照美国土壤PAEs化合物控制标准,我国土壤DBP含量的平均值全部超标,而DEHP含量未超标。

2.2 有/无温室大棚PAEs含量差异分析

温室大棚系统中DBP(DEHP)含量分别为:土壤0.457(0.733)mg·kg⁻¹,蔬菜1.508(1.831)mg·kg⁻¹;露天种植系统DBP(DEHP)含量分别为:土壤1.283(1.389)mg·kg⁻¹,蔬菜0.841(0.659)mg·kg⁻¹。通过对温室大棚与露天种植土壤和蔬菜PAEs含量对比分析发现,温室大棚种植体系中蔬菜的PAEs含量要高于露天种植(图2)。温室大棚棚膜和地膜在使用过程中,会向土壤和空气中持续释放PAEs,导致PAEs在土壤和蔬菜中的累积。研究发现,南京地区典型温室大棚蔬菜中Σ₆PAEs(DMP、DEP、DBP、BBP、DEHP、DnOP)累积量为0.79~7.3 mg·kg⁻¹,是露天蔬菜的2~3倍,其中DEHP浓度为4.3 mg·kg⁻¹,高于欧盟食品最高浓度限制值(1.5 mg·kg⁻¹)^[13],这主要由于棚膜中挥发出的PAEs在半封闭系统中的扩散能力受到一定的限制,温室大棚空气中PAEs浓度增加,同时土壤中PAEs挥发至温室大棚空气中,也会导致温室大棚蔬菜PAEs含量升高^[62]。统计分析表明,露天种植系统土壤PAEs含量要略高于温室大棚种植系统,这是因为在数据统计中有一些高值的出现,将露天土壤PAEs平均含量整体提升。



DBP-V和DEHP-V分别表示蔬菜中DBP和DEHP的含量,DBP-S和DEHP-S分别表示土壤中DBP和DEHP含量。下同
DBP-V and DEHP-V indicate the content of DBP and DEHP in vegetables, DBP-S and DEHP-S indicate the content of DBP and DEHP in soil, respectively. The same below

图1 不同地区土壤和蔬菜中的PAEs含量

Figure 1 The content of PAEs in soils and vegetables in different regions of China

2.3 不同蔬菜对PAEs的累积差异分析

统计发现,我国蔬菜中DBP和DEHP平均含量分别为 $1.271\pm1.732\text{ mg}\cdot\text{kg}^{-1}$ 和 $0.891\pm0.973\text{ mg}\cdot\text{kg}^{-1}$,检出率较高。对检索文献中所报道蔬菜中PAEs的含量进行统计(含量从大到小),以色阶的形式表示(图3)。通过降序排序发现,叶菜类蔬菜中生菜、葱花、芹菜、黄芽白、西洋菜、大葱、油麦菜、上海青、通菜、苦麦菜对PAEs富集较大(图3a),而瓜果类蔬菜(含谷物)

中冬瓜、蒲瓜、毛豆、茄子、梨、夏玉米籽粒、冬小麦籽粒、白薯对PAEs富集较大(图3b),均超过了全国蔬菜中 ΣPAEs 含量平均值($2.062\text{ mg}\cdot\text{kg}^{-1}$)。蔬菜生理结构特点,如叶片面积、叶片形状、根系类型、脂类含量和生长期是影响PAEs吸收的主要因素,此外,蔬菜的生长环境和气候条件也会影响PAEs浓度。研究显示,工业区蔬菜可食部分DEHP浓度为 $0.23\sim9.11\text{ mg}\cdot\text{kg}^{-1}$,高于非工业区蔬菜中DEHP浓度,其中叶菜类

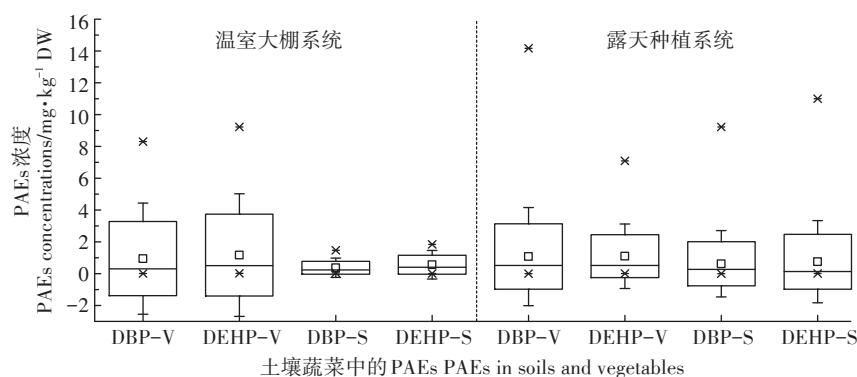


图2 温室大棚和露天种植系统中PAEs的含量

Figure 2 The content of PAEs in greenhouse vegetable system and uncovered farmland

| 蔬菜类型 | DBP-V | DEHP-V | ΣPAEs |
|------|-------|--------|---------------------|
| 生菜 | 2.62 | 3.72 | 6.34 |
| 葱花 | 0.58 | 2.67 | 3.25 |
| 芹菜 | 0.66 | 2.46 | 3.11 |
| 黄芽白 | 0.85 | 2.24 | 3.09 |
| 小麦 | 1.89 | 0.61 | 2.50 |
| 西洋菜 | 0.63 | 1.82 | 2.45 |
| 大葱 | 1.20 | 1.23 | 2.43 |
| 油麦菜 | 0.49 | 1.84 | 2.33 |
| 上海青 | 0.52 | 1.77 | 2.29 |
| 通菜 | 0.47 | 1.76 | 2.23 |
| 苦麦菜 | 0.49 | 1.60 | 2.09 |
| 韭菜 | 1.89 | 0.15 | 2.04 |
| 大白菜 | 0.24 | 1.65 | 1.89 |
| 小白菜 | 0.36 | 1.06 | 1.41 |
| 芥菜 | 0.34 | 0.92 | 1.26 |
| 番薯叶 | 0.45 | 0.75 | 1.20 |
| 花椰菜 | 0.55 | 0.54 | 1.10 |
| 芥兰 | 0.63 | 0.44 | 1.07 |
| 菜心 | 0.46 | 0.61 | 1.06 |
| 苋菜 | 0.22 | 0.76 | 0.98 |
| 白菜 | 0.61 | 0.23 | 0.84 |
| 蕹菜 | 0.43 | 0.21 | 0.64 |
| 葱 | 0 | 0.47 | 0.47 |
| 芥蓝 | 0 | 0.39 | 0.39 |
| 奶白菜 | 0 | 0.27 | 0.27 |

a. 叶菜类 Leaf vegetable

Red indicates high concentration, and blue means low concentration of PAEs, respectively

| 瓜果类PAEs含量 | DBP-V | DEHP-V | ΣPAEs |
|-----------|-------|--------|---------------------|
| 冬瓜 | 8.30 | 9.23 | 17.53 |
| 蒲瓜 | 3.23 | 1.11 | 4.34 |
| 毛豆 | 1.84 | 2.50 | 4.34 |
| 茄子 | 0.65 | 2.30 | 2.95 |
| 梨 | 0.88 | 1.74 | 2.62 |
| 夏玉米籽粒 | 1.31 | 1.20 | 2.51 |
| 冬小麦籽粒 | 0.90 | 1.50 | 2.40 |
| 白薯 | 0.77 | 1.40 | 2.17 |
| 茄瓜 | 1.47 | 0.46 | 1.93 |
| 丝瓜 | 1.31 | 0.31 | 1.62 |
| 水稻 | 0.96 | 0.42 | 1.38 |
| 黄瓜 | 0.70 | 0.60 | 1.30 |
| 番茄 | 0.78 | 0.50 | 1.28 |
| 豆角 | 1.04 | 0.22 | 1.26 |
| 辣椒 | 0.36 | 0.64 | 1.00 |
| 西葫芦 | 0.66 | 0.33 | 0.99 |
| 苦瓜 | 0.49 | 0.05 | 0.54 |
| 番薯 | 0 | 0.08 | 0.08 |

b. 瓜果类 Fruit and melon vegetable

红色越深表示浓度越大,蓝色越深表示浓度越小

Figure 3 The content of PAEs in different vegetables

DEHP含量最高,其次是果菜和根茎类蔬菜^[63]。研究表明叶菜类蔬菜中的生菜和果菜类蔬菜中的冬瓜PAEs含量最高(图3),可能与蔬菜的叶片大小及果实中脂类含量有关,使生菜和冬瓜较其他蔬菜具有对PAEs的积累能力^[25]。蔬菜对DBP和DEHP的积累量不同,这与蔬菜的种类和品种有关^[45]。根据已有文献数据,我们统计分析了不同类型蔬菜对DBP和DEHP的生物富集系数(表2)。通过降序排序发现蔬菜对PAEs的富集能力存在较大差异,如大白菜、茼蒿、莴苣等对DBP有较高的富集能力,而对DEHP富集能力不高;而芥蓝、奶白菜、上海青等对DEHP有较高的富集能力,而对DBP的富集能力不高。

2.4 日摄入量评价与人体健康风险

本次评估依据文献检索中的数据,暴露模型方程

表2 不同类型蔬菜对DBP和DEHP的富集系数

Table 2 Bioaccumulation factors of different types of vegetables

| 序号 No. | 蔬菜类型 Vegetable types | DBP | 序号 No. | 蔬菜类型 Vegetable types | DEHP |
|-----------|-------------------------|---------|-----------|-------------------------|----------|
| 1 | 大白菜 | 3.076 8 | 1 | 芥蓝 | 36.452 8 |
| 2 | 茼蒿 | 2.434 2 | 2 | 奶白菜 | 19.212 8 |
| 3 | 莴苣 | 1.814 5 | 3 | 上海青 | 16.510 1 |
| 4 | 蒜苗 | 1.556 6 | 4 | 菜心 | 14.728 8 |
| 5 | 菜心 | 1.454 5 | 5 | 番薯 | 7.045 0 |
| 6 | 菊花叶 | 1.393 8 | 6 | 菊花叶 | 6.097 5 |
| 7 | 芥蓝 | 1.230 8 | 7 | 芥菜 | 5.022 6 |
| 8 | 芥菜 | 1.133 3 | 8 | 蒜苗 | 3.683 8 |
| 9 | 芹菜 | 1.111 1 | 9 | 小白菜 | 2.132 0 |
| 10 | 菠菜 | 1.100 4 | 10 | 莴苣 | 1.834 6 |
| 11 | 上海青 | 0.727 3 | 11 | 菠菜 | 1.624 7 |
| 12 | 卷心菜 | 0.694 2 | 12 | 茼蒿 | 1.473 7 |
| 13 | 番薯 | 0.633 3 | 13 | 茄子 | 1.126 4 |
| 14 | 夏玉米 | 0.566 8 | 14 | 油菜 | 1.038 5 |
| 15 | 葱 | 0.548 4 | 15 | 夏玉米 | 0.978 0 |
| 16 | 小白菜 | 0.501 9 | 16 | 梨子 | 0.956 0 |
| 17 | 油菜 | 0.443 0 | 17 | 大白菜 | 0.931 6 |
| 18 | 夏小麦籽粒 | 0.426 7 | 18 | 冬小麦粒 | 0.824 2 |
| 19 | 白萝卜 | 0.425 5 | 19 | 芹菜 | 0.799 5 |
| 20 | 茄子 | 0.420 2 | 20 | 葱 | 0.781 0 |
| 21 | 大葱 | 0.390 9 | 21 | 甘薯 | 0.769 2 |
| 22 | 奶白菜 | 0.372 1 | 22 | 大葱 | 0.675 8 |
| 23 | 冬小麦粒 | 0.293 2 | 23 | 夏小麦籽粒 | 0.659 3 |
| 24 | 梨子 | 0.286 6 | 24 | 卷心菜 | 0.591 3 |
| 25 | 韭菜 | 0.255 6 | 25 | 白萝卜 | 0.363 3 |
| 26 | 甘薯 | 0.250 8 | 26 | 韭菜 | — |
| 27 | 山芋藤 | 0.063 5 | 27 | 山芋藤 | — |

注:富集系数=蔬菜中PAEs的含量/土壤中PAEs的含量。

如下^[64]:

$$DI = \frac{C \cdot IR}{bw} \cdot r_{\text{uptake}} \quad (1)$$

式中:DI是经口摄入PAEs的含量, $\mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$;C是PAEs在环境介质中的浓度, $\mu\text{g} \cdot \text{g}^{-1}$,饮用水 $\mu\text{g} \cdot \text{L}^{-1}$;IR是摄入量, $\text{g} \cdot \text{d}^{-1}$;bw是人的体重, kg ; r_{uptake} 是胃肠道从食物中吸收PAEs的速率。根据文献[51]的报道,将公式(1)修改为公式(2),以评价蔬菜中DBP和DEHP的暴露风险:

$$DI = \frac{C_{\text{DBP}} \cdot (1 - W) \cdot IR \cdot R}{bw} \cdot r_{\text{DBP}} + \frac{C_{\text{DEHP}} \cdot (1 - W) \cdot IR \cdot R}{bw} \cdot r_{\text{DEHP}} \quad (2)$$

式中:DI为经口摄入PAEs的含量, $\mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$; C_{DBP} 和 C_{DEHP} 分别表示DBP和DEHP在蔬菜中的浓度, $\text{mg} \cdot \text{kg}^{-1} \cdot \text{dw}$; W 表示蔬菜中的含水量,92.31%;IR是蔬菜日摄入量, $314.47 \text{ g} \cdot \text{d}^{-1}$; R 为大棚蔬菜占总蔬菜消费量的比例,冬季61%;bw为人的体重,幼儿(0.5~4岁)11.95 kg,儿童(5~11岁)26.15 kg,青少年(12~18岁)51.95 kg,成人(>18岁)65.56 kg; r_{DBP} 和 r_{DEHP} 表示胃肠道从食物中吸收DBP和DEHP的速率,DBP为0.685,DEHP为0.552。根据文献[65]所报道的数据进行计算,这些数据均依据现场测量、监测数据、调查和中国官方统计数据。

应用USEPA推荐的模型,对PAEs的非致癌风险和致癌风险进行评估。其中DBP为非致瘤物,而DEHP为致瘤物^[66]。蔬菜中DEHP的致瘤风险(CR)估算模型:

$$CR = q \cdot DI \quad (3)$$

式中: q 表示剂量-反应关系确定的致瘤斜率因子,对于DEHP,其值为 $0.014 \text{ mg} \cdot \text{kg}^{-1} \cdot \text{bw} \cdot \text{d}^{-1}$, $CR > 10^{-6}$ 表明存在致瘤风险。

蔬菜中DBP的非致瘤风险(HQ)估算模型:

$$HQ = DI / R/D \quad (3)$$

式中: R/D 表示日推荐剂量,对于DBP,其值为 $0.1 \text{ mg} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$, $HQ > 1$ 表明存在非致瘤风险。

目前,我国尚未制定农产品中PAEs含量限值,因此对PAEs污染水平和等级很难划分。根据欧洲经济共同体食品科学委员会的建议,每人PAEs的日摄入量不得超过 $0.3 \text{ mg} \cdot \text{kg}^{-1}$,USEPA指出,DBP的日摄入量不得超过 $100 \mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$,DEHP日摄入量不得超过 $20 \mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$ ^[69]。评价结果显示,幼儿较其他年龄组对DBP和DEHP有更高的暴露水平,其日摄入量分别为 $0.94 \mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$ 和 $0.87 \mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$ 。暴露水平

随着年龄的增加而减少,珠三角地区蔬菜PAEs幼儿暴露水平为 $1.526 \mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$,而成人暴露水平是 $0.278 \mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$;长江三角洲地区蔬菜PAEs幼儿暴露水平为 $2.084 \mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$,而成人暴露水平是 $0.389 \mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$;环渤海地区蔬菜PAEs幼儿暴露水平为 $2.297 \mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$,而成人暴露水平是 $0.419 \mu\text{g} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$ (图4)。因此,DBP和DEHP暴露水平在珠三角地区最低,而在环渤海地区最高。在任何年龄段DBP的暴露水平均要高于DEHP,但均低于USEPA规定的日摄入量^[9](图4)。非致癌风险和致癌风险分析结果显示,温室大棚蔬菜PAEs摄入风险大于露天种植蔬菜(表3),但均未超过其风险阈值。

3 温室蔬菜大棚PAEs综合调控措施

3.1 温室大棚设施改进

我们前期采集分析了塑料大棚空气中PAEs含量,结果显示,塑料大棚内空气中PAEs含量要显著高于大棚外空气(数据未发表)。增加温室大棚与外界

表3 人体健康风险评价
Table 3 Human health risk assessments

| PAEs | 不同年龄组成 Age groups | 大棚种植系统 Greenhouse vegetable system | 露天种植系统 Open field planting system |
|-----------------------|----------------------|--|---|
| DBP (非致癌风险) | 幼儿 | 6.37E-03 | 5.27E-03 |
| | 儿童 | 2.91E-03 | 2.41E-03 |
| | 青少年 | 1.46E-03 | 1.21E-03 |
| | 成人 | 1.16E-03 | 9.60E-04 |
| DEHP (致癌风险) | 幼儿 | 9.58E-03 | 7.75E-03 |
| | 儿童 | 4.38E-03 | 3.54E-03 |
| | 青少年 | 2.20E-03 | 1.78E-03 |
| | 成人 | 1.75E-03 | 1.41E-03 |
| HQ>1或CR> 10^{-6} | — | 0 | 0 |

的空气交换率是降低大棚空气内PAEs的有效途径,如增设大型通风设施或定期通风均可有效降低大棚空气中的PAEs,但通风会导致大棚中PAEs向周边扩散,造成一定程度的大气污染。为了解决该问题,我们构建了一套用于降解PAEs的空气生物过滤装置^[67],能有效去除蔬菜大棚空气中的PAEs。

为了解决塑料地膜对土壤的污染问题,液体地膜、生物可降解地膜、光降解地膜和光-生物降解地膜相继问世^[68-69],生物降解地膜,特别是淀粉基生物地膜,可以在自然条件下被微生物完全降解,对土壤无污染,具有广阔的前景^[36]。利用生物可降解地膜可提高玉米产量17.8%^[70],促进冬油菜养分吸收,减少土壤硝态氮累积^[71]。但若长期使用可降解地膜,可能会导致土壤质量下降,从而影响作物产量^[72]。

温室无土栽培以人工制造的作物根系环境取代了土壤环境^[73],可以有效地避免作物根系吸收土壤中包括PAEs在内的环境污染物。同时大棚无土栽培也可以避免设施栽培出现土壤连作障碍,有助于提高蔬菜品质^[53]。此外,将塑料大棚改造为玻璃大棚能够有效阻断PAEs从棚膜释放进入空气、土壤和蔬菜,但其改造费用较高。

3.2 温室大棚土壤改良

生物质炭作为土壤添加剂能够改善环境质量、提高土壤养分和作物产量^[74],其作用机制主要包括增加微生物生物量^[75]、促进腐殖质合成^[76]、增加土壤微量元素含量和有机污染物的生物利用性^[77-78]。添加生物质炭能够显著提高土壤对DEP的吸附能力;生物质炭对土壤吸附DEP的贡献率>80%,表明其对土壤DEP的吸附发挥主导作用^[79]。对种植作物的土壤分别施加0.5%、1.0%、5%的生物质炭,其产量分别增加了34.79%、67.41%、73.47%,施加生物质炭会减少作

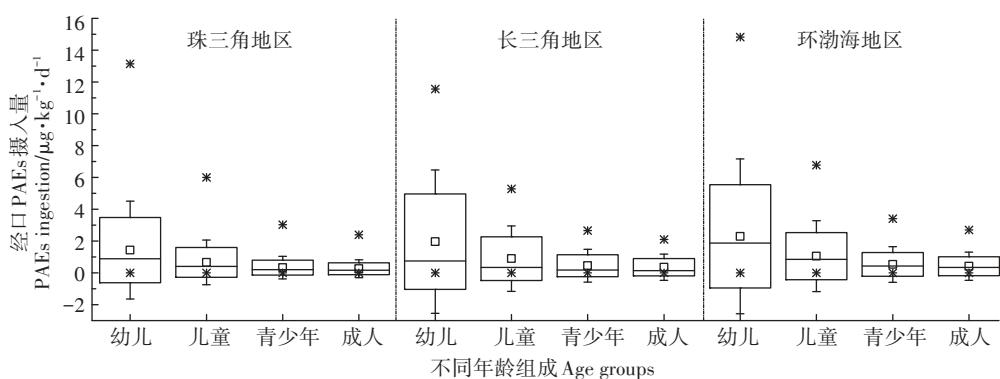


图4 不同地区蔬菜PAEs摄入量评价

Figure 4 Estimation of PAEs intake of vegetables in different regions of China

物对PAEs的吸收、迁移和富集^[80]。若将生物质炭施用于大棚土壤中,可能会降低PAEs从土壤向空气中挥发的能力,从而减少大棚空气和蔬菜中PAEs的含量。

近年来,微生物修复PAEs污染土壤已经被广泛研究和应用。PAEs降解菌能将PAEs作为唯一碳源,将其完全降解为CO₂和H₂O,且具有低成本、易操作和不引入二次污染等优点^[81]。前期,我们从土壤、城市污泥和垃圾填埋场中筛选了一系列高效降解PAEs的功能微生物,如微杆菌 *Microbacterium* sp. J-1^[82]、红球菌 *Rhodococcus* sp. 2G^[83]、普罗威登斯菌 *Providencia* sp. 2D^[84]等。其中菌株 *Microbacterium* sp. J-1能够降低根际土壤DEHP,降解率>97%,并且使接种蔬菜根部DEHP累积显著降低>70%^[85]。外源接种植物内生物能够直接去除植物体内的PAEs^[86]。目前可从研发含PAEs降解菌肥料入手,对土壤PAEs进行高效降解。

3.3 农艺调控措施

我们对农艺调控降低温室大棚系统PAEs的研究思路主要包括:(1)筛选PAEs低累积作物品种,如我们前期筛选获得了PAEs高/低吸收累积的作物品种(菜心、蕹菜、玉米等)及对有机污染物具有强降解作用的作物品种,获得了适合种植在PAEs中低程度污染土壤的蔬菜品种^[87-88];(2)PAEs低累积作物间作套种冬瓜,冬瓜是PAEs的“超富集”植物,其吸收能力要远强于其他蔬菜,研究表明冬瓜的叶、茎和果实均可从空气中吸收PAEs,单株冬瓜在暴露6周后能从空气中吸收超过700 mg·kg⁻¹的DEHP,其能够降低空气中65%~76%的PAEs^[89],因此,可将冬瓜与低累积品种蔬菜间作套种在温室大棚中,达到温室大棚PAEs边生产边修复的目的;(3)筛选能够促进根际微生物强力降解PAEs的作物;(4)作物-微生物联合修复PAEs。黄农1号玉米较对照(66.8%)在根际对DEHP具有更强的降解能力(87.5%),可达13.3~122 mg·盆⁻¹,植物体内DEHP也要少于其他作物^[75],这也为温室大棚PAEs污染控制与修复提供了一个可靠的依据。此外,向种植玉米的土壤中接种降解PAEs的菌株发现,其能够缓解PAEs对玉米的胁迫,保障玉米的生长^[90]。

4 结论与展望

温室大棚PAEs污染主要来自农膜/地膜、化肥农药和污水灌溉,PAEs在温室大棚土壤和蔬菜中具有较高的浓度和检出率。珠三角、长三角等4个地区土

壤和蔬菜中PAEs浓度具有差异性,这与区域农膜使用量有显著关系。温室大棚种植系统中蔬菜PAEs浓度要高于露天种植系统,而土壤PAEs含量分布则与之相反。蔬菜对PAEs的吸收积累与蔬菜的生理结构、种类、品种以及种植环境有关。蔬菜PAEs暴露水平在珠三角地区最低,环渤海地区最高。蔬菜中DBP的暴露水平要高于DEHP,但均低于USEPA规定的人均日摄入量。长期低剂量摄入含有PAEs的蔬菜仍然有一定的健康风险。目前,对温室大棚系统重金属污染修复已有较多报道,而对PAEs等有机污染问题研究较少。因此,需要加强以下几方面的研究:

(1)增加设施农业有机污染物调查种类。温室大棚中的有机污染物不仅有PAEs,还包括农药、抗生素和全氟化合物等,这些污染物对人体具有较大的健康风险,应加强这些有机污染物的调查研究。

(2)温室大棚种植系统中土壤-空气-蔬菜体系PAEs等有机污染物的环境行为研究。研究PAEs有机污染物在温室大棚中的环境行为,如土壤中的吸附-解吸、水解、土壤-空气界面交换率、微生物降解、植物吸收累积等过程,为控制温室大棚PAEs等有机污染物提供科学依据。

(3)建立一套适合温室大棚种植系统的PAEs污染修复技术。温室大棚常年处于封闭或半封闭状态,挥发性/半挥发性有机污染物不易扩散,导致温室大棚蔬菜较露天种植蔬菜更容易吸收有机污染物,因此改善大棚通风条件、增设通风设施是关键。此外,水肥管理、农药施用、低累积作物筛选、土壤改良剂也是温室大棚PAEs等有机污染物控制技术研究的方向。

(4)加强管理。为了降低温室大棚PAEs等有机污染,需要不同利益相关者(如政府、企业、合作社、经销商、种植户等)的协同努力,并对温室大棚运营进行生命周期评估,同时依靠环境保护法律法规的有效执行。制定与温室大棚有关的环境质量标准和先进的控制技术,以保障温室大棚蔬菜的质量安全。

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